



Development of Numerical Model for Colloid-Facilitated Transport of Multiple Members of a Radionuclide Decay Chain

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. Introduction / Deep geological disposal for high-level waste

- Nuclear power generation provides clean energy, but the disadvantage is that it produces radioactive waste.
- Spent nuclear fuel has high level of radioactivity and a long half-life. This type of radioactive waste is called high-level waste (HLW).
- The concept of dealing with HLW is to isolate HLW from the biosphere until the concentration of HLW decays to a harmless level.

> For the final disposal of HLW, the main method is using deep geological disposal.

- 1. Multi-barrier system
- 2. Low-permeability hard rock formation
- 3. 500 meters or more below the surface

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. Introduction / Transport of radionuclides in the fracture-matrix system

- The fractures may appear due to the mechanical stress created during the construction of deep geological disposal facilities and thermal and radiation effects due to the presence of HLW.
- The mechanism by which radionuclides in a deep geological disposal could return to the biosphere is the movement of groundwater through the fractures.





- > Colloids are present in groundwater, can sorb radionuclides and transport in groundwater.
- > Colloids are substances with sizes ranging from 1 nm to 10 μm .
- Many substances can exist in colloidal form, including fine clay particles, humic substances, bacteria, radionuclides, etc.



Matrix
Fracture surface
Fracture
Fracture surface
Matrix
(Mobile colloid) Immobile colloid

Introduction / Feature of colloid

Why do colloids facilitate transport of radionuclides?

When radionuclides are sorbed onto the colloids, they can no longer diffuse into the matrix. So, the retardation effect of matrix diffusion is lost. Therefore, the distance that the radionuclides move will increase.





Radionuclides with colloids

Radionuclide in groundwater in fracture

Radionuclide sorption onto mobile colloid



. Introduction / Colloid-facilitated the transport of a radionuclide decay chain







Immobile colloid

.Introduction / Literature review

- Traditional two-phase methods contain only matrix and groundwater in fractures, which don't consider the role of colloids, therefore tend to underestimate the transport of contaminants. (Tang et al., 1981; da Silveira et al., 2013; Bagalkot and Kumar, 2015)
- Field-scale and laboratory-scale observations have shown that colloids in groundwater can act as a third phase which can sorb contaminants. (McCarthy and Zachara, 1989)
- At a liquid waste disposal site in New Mexico, previous laboratory studies indicated that contaminants should be retained in the top few millimeters of the soil profile. But in the end, it was found in-situ that it transported over 30 m. Because the effect of colloids is ignored. (Nyhan et al., 1985)

. Introduction / Literature review

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	Numerical solution	Single radionuclide	Radionuclide decay chain	Colloid concentration
Baek and Pitt Jr, 1996	0	Ο		constant
Kheirabadi et al., 2016	0	Ο		constant
Chopra et al., 2016	0		О	constant



From Baek and Pitt Jr, 1996

Ignoring the temporal and spatial variation of colloid concentration will affect the correctness of exploring the influence of colloids on the transport of multiple members of a radionuclide decay chain.



To develop a numerical model for colloid-facilitated transport of multiple members of a radionuclide decay chain which can consider the temporal and spatial variation of colloid concentration.



Colloid transport in fracture 1.

$$\frac{\partial C_c(x,t)}{\partial t} = D_f \frac{\partial^2 C_c(x,t)}{\partial x^2} + v_f \frac{\partial C_c(x,t)}{\partial x} + \frac{r_s}{b}$$

Dispersion Advection





Mobile colloid



Immobile colloid

$C_c(x,t)$	The concentration of colloids in the water in the fracture (ML^{-3})
x	The distance along the fracture from the source (L)
t	The time (T)
D_f	The dispersion coefficient of water in the fracture (L^2T^{-1})
v_f	The velocity of water in the fracture (LT^{-1})
r_s	The rate of colloid capture on the fracture surfaces (ML^{-2})
b	The half fracture-aperture (<i>L</i>)

. Mathematical model / Governing equation

2. Radionuclides transport in matrix



 $\frac{\text{Decay}}{\text{chain}} \left\{ \begin{array}{c} 1 \\ 2 \\ 3 \\ 4 \end{array} \right\}$

Radionuclides diffusion into matrix

. Mathematical model / Governing equation

3. Colloid-facilitated radionuclides transport in fracture



2 3 4 Radionuclide sorption onto mobile colloid

chain

• **4** Radionuclide sorption onto immobile colloid

Mathematical model / Governing equation

4. The rate of diffusion from the fracture into the matrix

 $q_k(x,t) = -\phi_m D_m \frac{\partial C_{m,k}(x,z,t)}{\partial z} \bigg|_{z=1}$



$q_k(x,t)$	The diffusion flux from fracture water into rock matrix of the <i>k</i> th radionuclide $(ML^{-2}T^{-1})$
ϕ_m	The porosity of rock matrix (–)
D _m	The diffusion coefficient in the pore water (L^2T^{-1})
$C_{m,k}(x,z,t)$	The concentration of the <i>k</i> th radionuclide in water in rock pores (ML^{-3})
b	The half fracture-aperture (L)

Decay chain **1 2 3 4** Radionuclide in groundwater in fracture **1 2 3 4** Radionuclides diffusion into matrix

. Mathematical model / Conceptual model



. Mathematical model / Flow chart for this research





. Results and discussion / Verification

Case 1 : Single radionuclide without colloid



. Results and discussion / Verification

Case 2 : Single radionuclide with colloid



Results and discussion / Verification

> Case 3 : A three-member decay chain of ^{238}U without colloid



Results and discussion / Parameter for sensitivity analysis

Parameter	Value	Parameter	Value	Parameter	Value
v _f (m/year)	10	<i>L</i> (<i>m</i>)	1000	$K_1(m)$	0
$D_f (m^2/year)$	100	W (m)	5	<i>K</i> ₂ (<i>m</i>)	0
$D_m (m^2/year)$	0.01	φ(-)	0.01	$K_3 (m^3/kg)$	100
$R_f(-)$	1	t (year)	100	$K_4 (m^3/kg)$	100
$R_m(-)$	100	λ (year ⁻¹)	3.24E-07	$C_c (kg/m^3)$	0.1
<i>b</i> (<i>m</i>)	0.005	$C_{f,0} \left(kg/m^3 \right)$	1	(Baek and Pi	tt Jr, 1996)

 K_1 is the distribution coefficient of colloids with fracture surface.

 K_2 is the distribution coefficient of radionuclides from water in fracture to fracture surface.

 K_3 is the distribution coefficient of radionuclides with mobile colloid.

 K_4 is the distribution coefficient of radionuclides with immobile colloid.

Distribution coefficient (K_3) is small.

Distribution coefficient (K_3) is large.





Parameter	Value	Parameter	²³⁸ U	²³⁴ Th	^{234}U	²³⁰ Th	²²⁶ Ra
$v_f (myear^{-1})$	50	λ (year ⁻¹)	1.54E-10	10.5	2.80E-6	8.66E-6	4.33E-4
$D_f (m^2 y ear^{-1})$	1	$R_m(-)$	27001	54001	27001	54001	54001
$D_m (m^2 y ear^{-1})$	0.005	$C_{f,0} (kgm^{-3})$	1	-	-	-	-
b (m)	0.001	$K_1(m)$	2.18E-4	4.37E-4	2.18E-4	4.37E-4	4.37E-4
L (m)	500	$K_2(m)$	0	0	0	0	0
W (m)	5	$K_3 (m^3/kg)$	40	40	40	40	40
φ (-)	0.01	$K_4 (m^3/kg)$	40	40	40	40	40
t (year)	5000	(Chopra et al., 2016)					

. Results and discussion / Colloid-facilitated the transport



Results and discussion / Effect of colloid concentration (C_c)

✓ Considering the colloid concentration as constant overestimates the transport of radionuclides.



. Results and discussion / Effect of distribution coefficient (K_3)

✓ Higher values of K_3 enhanced the transport of radionuclides.

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Results and discussion / Effect of porosity of matrix (ϕ)

 \checkmark As the value of ϕ increases, the transport of radionuclides is reduced.







- This model can simulate the colloid-facilitated transport of multiple members of a radionuclide decay chain which can consider the temporal and spatial variation of colloid concentration.
- > The presence of colloids can facilitate transport of radionuclides in fracture.
- Considering that the colloid concentration is constant will overestimate the transport of radionuclides.
- This model can be a useful tool for safety assessment studies of deep geological disposal for HLW.

Thank you for your attention